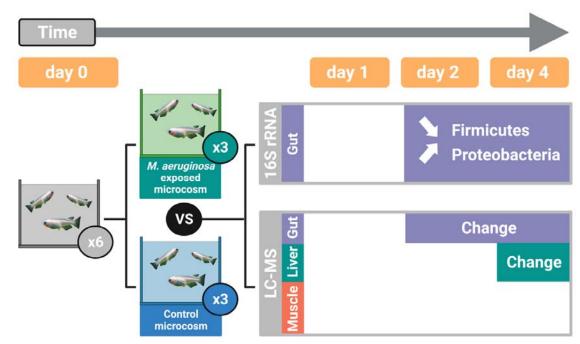
- 1 Gut microbiota and holobiont metabolome composition of the Medaka fish (Oryzias
- 2 *latipes*) are affected by a short exposure to the cyanobacterium *Microcystis aeruginosa*
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- 4
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9 Graphical abstract

1011 Highlights:

- A 2-day exposure to a simulated *M. aeruginosa* bloom is enough to sharply decrease
 the Firmicute/Proteobacteria ratio in the gut of *Oryzias latipes* fish.
- The exposure induced changes in metabolome composition after 2 days in the gut and
 4 days in the liver.
- The gut bacterial microbiota response occurred faster than metabolome's; we
 hypothesize that changes in gut microbiota may drive the gut metabolome
 compositional changes.

19 Abstract

20 Blooms of toxic cyanobacteria are a common stress encountered by aquatic fauna. 21 Evidence indicates that long-lasting blooms affect fauna-associated microbiota. Because of 22 their multiple roles, host-associated microbes are nowadays considered relevant to 23 ecotoxicology, yet the respective timing of microbiota versus functional changes in holobionts 24 response needs to be clarified. The response of gut microbiota and holobiont's metabolome to 25 exposure to a dense culture of *Microcystis aeruginosa* was investigated as a microcosm-26 simulated bloom in the model fish species Oryzias latipes (medaka). Both gut microbiota and 27 gut metabolome displayed significant composition changes after only 2 days of exposure. A 28 dominant symbiont, member of the Firmicutes, plummeted whereas various genera of 29 Proteobacteria and Actinobacteriota increased in relative abundance. Changes in microbiota 30 composition occurred earlier and faster compared to metabolome composition, suggesting that 31 the microbiota drives the holobiont's response. Liver and muscle metabolome were much less 32 affected than guts, supporting that gut and associated microbiota are in the front row upon 33 exposure. This study highlights that even short cyanobacterial blooms, that are increasingly 34 frequent, trigger changes in microbiota composition and holobiont metabolome. It emphasizes 35 the relevance of multi-omics approaches to explore organism's response to an 36 ecotoxicological stress.

37 Keywords: Microbiota; Ecotoxicology; Multi-omics; Time-series; Cyanobacteria; Holobiont

38 1. Introduction

39 One of the most common yet intense stress encountered by fauna in ponds and lakes is 40 the occurrence of algal bloom involving noxious cyanobacteria, which are increasing in 41 relation to anthropogenic effects¹. Adverse effects of blooms, ranging from specific organ 42 toxicity (e.g. hepatotoxicity and cardiotoxicity) to reproduction alteration, are increasingly documented in various animal models^{2–5}. On the other hand, effects of blooms on associated 43 44 microbiota are barely known. Beyond their role in holobiont development and functioning 45 (nutrition, immunity, protection, behavior...), host-associated microbial communities are now considered fully relevant to ecotoxicology $^{6-9}$. Indeed, communities are often located at the 46 47 interface between a host and its environment. They react to and interact with contaminants, with outcomes ranging from inactivation to potentialization¹⁰. Because of their relative 48 49 limited diversity of gut microorganisms, notably compared to endotherms, and of their 50 patrimonial and economic value, teleost fish are particularly relevant vertebrate models for

aquatic microbiome-aware ecotoxicology⁷. Among these, the medaka *Oryzias latipes* and the
zebrafish *Danio rerio* are commonly employed for ecotoxicological studies.

53 Recently, 28-day exposure experiments conducted in microcosms on the medaka 54 revealed that extracts of Microcystis aeruginosa containing microcystin-LR (MC-LR) among other bioactive metabolites could trigger dysbiosis in gut microbiota¹¹. Since then, similar 55 effects were shown in the zebrafish^{12,13}. A recent study used microcosm-simulated *M*. 56 57 aeruginosa blooms by exposing fish to high density of a cyanobacterial culture. Results 58 revealed dose-dependent alterations of the composition of medaka gut bacterial communities 59 and the holobiont metabolome, including a sharp decrease in abundance of a significant Firmicutes symbiont¹⁴. Altogether, these studies demonstrate that long exposures to 60 61 cyanobacteria, together with their subsequent metabolites affect fish microbiota, and at least 62 one of those studies also suggests an alteration of holobiont functions correlated with changes in microbiota composition¹⁴. In nature though, cvanobacterial blooms are often shorter than 63 64 the typical 20-28 days exposure used in ecotoxicological studies and documenting the early 65 dynamics of changes is the key to understanding holobiont response in natural relevant 66 settings.

67 In this study, we investigated the early response of the medaka gut microbiota and 68 organ metabolomes (gut, liver and muscle) to a 4-day exposure to high yet environmentally 69 relevant density of *M. aeruginosa*, the main species responsible for blooms in temperate lakes 70 and ponds. Compositions of the gut bacterial communities and gut, liver and muscle 71 metabolomes are monitored using 16S rRNA gene sequencing and LC-MS metabolomics, 72 respectively. Alteration dynamics of these compartments are then compared to highlight the 73 key role of gut microbiota in the holobiont functional response. This study is the first attempt 74 to address the initial short-term response of fish gut microbiota and holobiont metabolome to 75 a simulated cyanobacterial bloom.

- 76
- 77 2. Material and methods
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79 2.1 Exposure experiments

Experiments were performed in six aquaria (10-liters each, 3 control and 3 exposed, assigned randomly), each containing 15 specimens of 6-months old adult male medaka. Aquaria were stabilized for one month and fishes acclimatized for one week prior to exposure. At day 0, nine fish in total were sampled randomly across the six aquaria and 25 mL of water 84 from each aquarium were pooled as controls. Fish were then exposed for 4 days to water, or 85 water containing a strong but environmentally relevant density of the live non-axenic monoclonal *M. aeruginosa* strain PMC 728.11¹⁵ of the Paris Museum Collection¹⁶ to simulate a 86 bloom (100 μ g.L⁻¹ Chl a). The strain was cultured accordingly to anterior publication^{11,14} 87 88 (details in the supplementary information). M. aeruginosa concentrations were estimated after 89 performing Chlorophyll a extractions and absorbance measurements as a proxy using a spectrophotometer¹⁷ (Cary 60 UV-Vis, Agilent). Culture was sampled for DNA (1 mL) and 90 91 metabolome (50 mL) analyses on days 0 and 2. The *M. aeruginosa* level in exposed aquaria 92 was adjusted on days 0 and 2 to maintain exposure level. Three fish per aquarium were 93 sampled on days 1 and 2 and four per aquarium at day 4 (3 in one aquarium), as well as water 94 samples.

95 Water parameters were monitored on days 0, 2 and 4 (pH, temperature, conductivity, 96 nitrates and nitrites), feces were removed daily by aspiration, and half of the water was 97 replaced with freshwater (2/3 osmosis (RiOs 5, Merck Millipore) and 1/3 filtered) containing 98 or not, adjusted amount of Microcystis cells. Fish were exposed to constant temperature (25.4 ± 1 °C), pH (7.61 ± 0.1) and conductivity (246 $\pm 12.4 \ \mu \text{S.cm}^{-1}$), to low levels of nitrates and 99 100 nitrites (Table S1) to a 12h:12h light/dark cycle. They were fed twice daily (~3-5% of the fish 101 biomass per day) with Nutra HP 0.3 (Crude protein 57, Crude fat 17, N.F.E 7.5, Ash 10, 102 Crude fiber 0.5, Phosphorus 1.7, Vitamins A, D3, E; Skretting, Norway). Total microcystines 103 (MCs) levels were quantified in duplicates on days 0 and 4 (details in the supplementary 104 information).

105

106 2.2 Metabolites extraction and characterization

107 Metabolites were extracted from flash-frozen dissected medaka guts, livers, muscles, 108 and from lyophilized *M. aeruginosa* cultures. Mechanical extraction (GLH850 OMNI; 25 000 109 r.min⁻¹; 30s) followed by sonication (Sonics Vibra-Cell VCX 13; 60% amplitude; 30s) were 110 performed on weighted samples suspended in the extraction solvent (75-25% UHPLC 111 methanol-water, 1 mL per 100 mg of tissue or per 10 mg of lyophilized culture, on ice). After 112 centrifugation (10 min; 4 °C; 15,300 g), gut pellets were dried and used for subsequent DNA 113 extraction¹⁴.

Supernatants containing metabolite extracts were analyzed by Ultra high-performance liquid chromatography (UHPLC, Elute Bruker) using a Polar Avance II 2,5 pore C18 column (300 μ L.min⁻¹, Thermo) coupled with a high-resolution mass spectrometer (ESI-Qq-TOF, Compact Bruker) at 2 Hz speed on positive simple MS mode. Feature peak lists were

generated from MS spectra within a retention time window of 1-15 minutes and a filtering of 5000 counts using MetaboScape 4.0 software (Bruker). The peak lists consisted of the areaunder-the-peaks of extracted analytes from the three tissues (medaka's gut, liver and muscles) sampled on days 0, 1, 2, 4 and the *M. aeruginosa* lyophilized culture. A log transformation was applied to metabolomics datasets. Principal Component Analysis (PCA) were performed using the mixOmics¹⁸ (v6.14.1) R package.

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2.3 DNA extraction, sequencing and analysis of the V4-V5 region of bacterial 16S rRNA

126 DNA was extracted from the pellets of the guts and *M. aeruginosa* culture, and from 127 $0.22 \ \mu m$ filters for aquarium water. Extractions were performed using the ZymoBIOMICS 128 DNA Mini-prep kit with a FastPrep 5G beat beater disruption (DNA Matrix; 4x30s; 6m.s⁻¹) 129 following manufacturer's instructions. An extraction-blank control sample was also performed. The V4-V5 region of the 16S rRNA encoding gene was amplified using primers 130 515R and 926F¹⁹ and sequenced on an Illumina MiSeq 250x2 bp platform (Biomnigene, 131 132 Besançon, France). Reads were deposited into the Sequence Read Archive (SRA) database 133 (accession number PRJNA836730 (samples SRR19170691-SRR19170767; Table S2).

Sequence analysis including primer removal and quality control was performed using the QIIME2-2021.2 pipeline²⁰. Forward and reverse reads were trimmed at 250 and 200 bp, respectively. Amplicon Sequence Variants (ASVs) were obtained with DADA2 (default parameters) and affiliated with the SILVA 138-99 database. Diversity metrics were computed with the phyloseq²¹ (v1.34.0) R package. Statistical analyses were performed using R packages vegan²² (v2.5-7) and RVAideMemoire²³ (v0.9-79). All values are displayed as median ±standard deviation.

141

142 2.4 MEBA analysis

A Multivariate Empirical BayesianAnalysis^{24,25} (MEBA) was performed to discriminate differentially abundant taxa through time and between treatments. Relative abundance tables at the Phylum and Genus taxonomic levels were obtained from phyloseq and analyzed using the MEBA plugin from Metaboanalyst 5.0 (https://www.metaboanalyst.ca), with no data transformation and all by-default parameters. Taxa with a MEBA T² score superior to 5 were consider discriminant and further statistically analyzed.

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150 2.5 ASV-metabolites correlative networks

A correlative network analysis was performed with DIABLO²⁶, a multi-omics framework created by the MixOmics team¹⁸, to analyze putative correlated dynamics between gut microbiota ASVs and gut metabolites. Briefly, the *block.plsda()* function performs a Pattern Latent Structure Discriminant Analysis which provides a reduced-dimension space with covariance-maximizing axes for each dataset (named "block"). A correlation score for the given blocks was computed with the *plotDiablo()* function.

157

158 **2.6** Comparative dynamics of microbiota and metabolome

159 The composition change dynamics in microbiota versus metabolome were investigated 160 by comparing trajectories of centroids using a newly developed method: MOTA (Multivariate 161 Omics Trajectory Analysis). Microbiota and metabolome datasets were log transformed and 162 analyzed separately but in a similar way. Distances between PCAs' centroid coordinates were 163 computed to create a trajectory between each sampling day for the two treatments and 164 displayed as the fraction of the total length achieved at each day (from 0% to 100% between 165 day 0 to 4). Trajectories were plotted for the 16S rRNA versus metabolomic data, allowing 166 comparison of their respective dynamics in the two treatments (details in the supplementary 167 information).

168

169 **3. Results and discussion**

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171 Concentrations of *M. aeruginosa* varied between 31 and 126 μ g.L⁻¹ Chl *a* in exposed 172 aquaria, well above values recently documented to alter microbiota composition after a 28-173 day exposure¹⁴, and are thus appropriate to investigate short-term effect and early response of 174 the holobiont. In these aquaria, MCs levels were high at day 1 and increased at day 4 (12 ±5 175 to 21 ±9 μ g eq. L⁻¹ MC-LR, respectively; Table S1). These conditions, above values reported 176 from extracts used previously¹¹, are expected to produce major changes in a 14-day exposure 177 and are thus appropriate to evaluate short term effects.

178

179 3.1 Rapid shift in community compositions during exposure to Microcystis aeruginosa

Analysis of bacterial community compositions based on the V4-V5 region of 16S rRNA-encoding gene from 67 fish guts, 7 waters, 2 *M. aeruginosa* cultures and 1 extraction blank samples yielded 3,702,081 raw reads, of which 73.4% were retained after sequence assembly, denoising and chimera removal. Samples displayed 21,134 to 59,009 reads (31,638 \pm 7,000), except the extraction blank (520 reads; Table S3). After taxonomic assignment and

185 removal of eukaryotes, mitochondria and chloroplasts, sequences clustered into 936 ASVs. Of 186 these, 856 were considered as abundant (they represented at least 1% of reads in at least 1 187 sample). Rarefaction curves reached saturation confirming that our sequencing effort was 188 sufficient to describe most of the bacterial diversity (not shown). Alpha diversity indices did 189 not indicate major changes among microcosm compartments (culture, water and fish guts) or 190 between dates and treatments (Fig S1; Table S4). Although bacterial ASVs richness was 191 higher in fish guts than other compartments and bacterial ASVs richness and evenness 192 increased during the experiment in both exposed and unexposed fish guts, differences were 193 not statistically significant.

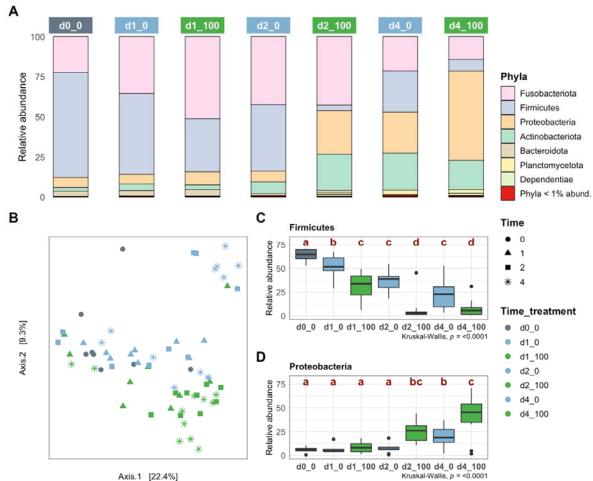
194 Community compositions were significantly different in fish guts compared to water 195 and culture samples (Unweighted UniFrac; Permanova p < 0.05; Table S5). Among fish guts, 196 composition differences were significant (Unweighted UniFrac; Permanova p < 0.05; Table 197 S5). Visually, samples from different time points are scattered along the first PCoA axis and 198 the two treatments are well separated on the second axis of the PCoA (Fig 1B). Indeed, the 199 microbiota of fish exposed to high density of *M. aeruginosa*, or non-exposed, displayed 200 compositions different from d0 at both d2 and d4. In addition, significant differences between 201 the two treatments were observed at d2 and d4 (Unweighted UniFrac; Pairwise.Permanova 202 p < 0.05; Table S5), indicating that gut community compositions were affected by exposure 203 duration (day 0 versus 2 and 4) and treatment (water versus *M. aeruginosa* exposed). Intra-204 group variances were not significantly different, allowing group comparison (Permdisp 205 *p*>0.05; Table S5).

206 Firmicutes was the dominant phylum in the fish gut bacterial communities at day 0 207 (64.7 $\%\pm 6$ of reads; Fig. 1A) and was discriminant among time and treatment (MEBA T² 208 score>5; Table S6). In *M. aeruginosa*-exposed fish, Firmicutes decreased significantly 209 between day 0 and 1 (64.7 \pm 6% to 33.7 \pm 14.4%; Wilcoxon p<0.05; Fig 1C; Table S6), then 210 day 2 (to 2.91 \pm 14.24 %; Wilcoxon *p*<0.05; Table S6) after when then remained stable at day 211 4 (Wilcoxon p>0.05; Table S6). Firmicutes decreased less dramatically in unexposed fish. 212 Significant decreases were observed on days 2 and 4 (38.8 $\pm 12.3\%$ and 22.78 $\pm 16.49\%$ 213 respectively; Wilcoxon p < 0.05; Table S6). The gap between exposed and unexposed fish on 214 days 2 and 4 was significant, supporting that the treatment itself further decreased Firmicutes 215 abundance on top of the effect observed in unexposed fish. Among discriminant taxa, this 216 decreasing pattern was only observed with Firmicutes and its main genus ZOR0006 (Table 217 S6). A similar decrease of ZOR0006 was observed after 28 days exposure to even moderate levels of *M. aeruginosa*¹⁴. Based on its genome content¹⁴, *ZOR0006*, a dominant resident in 218

guts of healthy medaka¹⁴, can perform lactate pyruvate interconversion, a function essential to 219 220 the gut homeostasis as lactate generally inhibits the growth of most pathogens while pyruvate stimulates it^{27–29}. Lactate can also be involved in the repair of the gut epithelium, yet high 221 levels can also be associated with inflammatory bowel disease in humans^{30,31}. The decrease of 222 Firmicutes has been considered an indicator of dysbiosis²⁴, here suggesting that exposure to 223 *M. aeruginosa* triggers a strong dysbiosis as early as 2 days. The moderate decrease perceived 224 225 in unexposed specimens could be a consequence of the stress associated with specimen handling³². 226

227

Contrariwise, Proteobacteria were also discriminant (MEBA T² score>5), but their 228 abundances increased significantly from day 2 in *M. aeruginosa*-exposed fish. It was even 229 230 more intense at day 4 (18.70 \pm 11.34 to 45.21 \pm 21.65%; Fig 1D; Wilcoxon p<0.05; Table S6), 231 while it was significant only at day 4 for unexposed fish (Wilcoxon p < 0.05; Table S6). As for 232 the decreasing pattern, the gap between the unexposed and exposed samples on days 2 and 4 233 was significant (Wilcoxon p < 0.05; Table S6), indicating that the treatment influenced the 234 abundance of the Proteobacteria on top of the effect observed in unexposed fish. Four 235 discriminant genera (Xanthobacter, Reyranella and Devosia belonging to Proteobacteria and Nocardioides, phylum Actinobacteriota) displayed a similar pattern (MEBA T² score>5: 236 Table S6). Other discriminant taxa abundances were not significantly altered between 237 2 238 treatments both 4 (Table on days and S6).



rig. 1: Composition and diversity of fish gut bacterial microbiota among time and treatment.

241 A: Gut microbiota composition at the phylum level (median values of samples belonging the same 242 time treatment group). B: Principal Coordinates Analysis (Unweighted UniFrac distance). C-D: 243 Relative abundance variations over time in both treatments, observed in Firmicutes (C) and 244 Proteobacteria (D); letters refer to Benjamini-Hochberg (BH) adjusted Wilcoxon post-hoc test 245 significance.

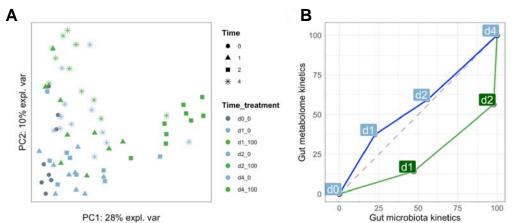
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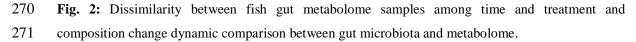
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247 3.2 Exposure to Microcystis aeruginosa induces changes mostly in gut metabolome 248 *composition*

249 Mass spectrometry distinguishes 921, 2,521 and 4,190 metabolites present in gut, 250 liver, and muscle samples, respectively. The PCA analysis visually separates gut 251 metabolomes of fish exposed to *M. aeruginosa* for 2 and 4 days, away from other samples 252 (Fig. 2A). Metabolome compositions were significantly altered on days 2 and 4 for exposed 253 fish and, to a lesser extent, unexposed fish (Permanova p < 0.05; Table S5). The metabolite 254 compositions between treatments on days 2 and 4 was different (Permanova p < 0.05; Table

255 S5), suggesting an additional effect of the treatment to the changes observed in unexposed 256 fish. This assumption is further supported by the observation that 574 of the 916 observed in 257 gut metabolites displayed differential abundances among dates in exposed specimens versus 258 only 152 in unexposed fish (Anova p < 0.05; Fig S2A, B). Liver metabolome composition is 259 significantly different between day 4 and other dates, and at this day between treatments, 260 suggesting an effect of both the duration of the experiment and the *M. aeruginosa* exposure 261 after 4 days (Permanova p < 0.05; Table S5). Intra-group variances were not significantly 262 different, allowing group comparison (Permdisp p>0.05; Table S5). Finally, composition of 263 muscle metabolomes did not display any significant variation (Permanova p > 0.05; Table S5). 264 Thus, the gut functional status appears altered earlier than the livers, while the muscles 265 functional status is not affected. This is not surprising given that cyanobacteria enter fish 266 through oral ingestion and their bio-active metabolites transfer through the intestine, which 267 could buffer their effects before other organs are affected, and thus protect the host during short blooms^{4,33}. 268





272 A: Principal Component Analysis of gut metabolite compositions of exposed and unexposed fish over 273 time. B: Trajectory of the composition change in microbiota versus metabolome expressed as 274 percentages of the total trajectory achieved on days 0, 1, 2 and 4 on each axis.

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3.3 Do changes in gut bacterial communities drive functional changes in the holobiont?

277 In unexposed fish, the percentages of the total distance achieved by gut metabolome and 278 microbiota compositions are similar on days 1 (22 and 37%) and 2 (56 and 59%), resulting in 279 a trajectory that is close to the 1:1 diagonal (Fig. 2B), possibly reflecting an overall drift of the system due to the aforementioned handling stress 32 . The shape of the trajectory is different in 280

281 exposed fish. Indeed, the gut metabolome achieved 14% of the total distance at day 1 versus 282 47% for the microbiota. At day 2, cumulative values are 56% for the metabolome and 98% 283 for the microbiota, suggesting that most of the change ultimately observed in the microbiota 284 composition is achieved after only 2 days. This holds true also when accounting for a higher 285 percentage of the total explained variance (Fig S3). The trajectory thus shifts towards the gut 286 microbiota axis for exposed samples, suggesting a faster response of the gut microbiota 287 compared to the gut metabolome. The gut microbiota bacterial ASVs and gut metabolites 288 datasets were well correlated at both day 2 and 4 (correlation score: 0.91). These results 289 suggests that changes in microbiota might be preceding, and possibly driving observed 290 metabolome changes, a hypothesis congruent with the localization of the gut microbes at the 291 interface between host and its environment (digestive lumen), and thus their role as a primary barrier to contaminants^{7,10}. Although it is tempting to infer causality here, further 292 293 confirmation is needed since the gut metabolome contains both host and microbiota-derived 294 metabolites, which could amplify the metabolome variations when bacterial communities 295 change a lot.

4. Conclusion

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298 Fish gut bacterial community composition and metabolome are affected within the 299 first two days upon exposure to *M. aeruginosa*. Thus, even short cyanobacterial blooms can 300 trigger drastic variations in bacterial phyla abundances, with consequences for the functioning 301 of the gut. This way, iterative short bloom events could lead to major shifts in gut microbiota 302 compositions and its associated functions, for example by progressively eliminating some 303 sensitive resident symbiont lineages, exemplified here by the Firmicutes ZOR0006. Although sex-specific responses to perturbations have been observed in ecotoxicological³⁴ and gut 304 305 microbiota³⁵ studies, only male fish were used here to first test the existence of a short-term 306 response. Further experiments are now needed to address alterations in female fish. This study 307 highlights the relevance of time-series exposure experiments and complementarity of omics 308 approaches, using metabolomics as a proxy of the integrated functional response, to address 309 short and long-term responses of holobionts to ecotoxicological stress, and the respective 310 roles played by host and microbes.

311

312 Author's contributions

313 A.G., B.M. and S.D. conceived the study. A.G., C.D. and B.M. and S.D. conceived the

314 experiment. P.F., A.G. and C.D. conducted the experiment. B.M. and S.D. took part in the

315	experiment. P.F.	., A.G. and C.I	D. conducted	molecular dat	a processing.	P.F. B.M.,	and S.D.

- analyzed data. P.F., B.M. and S.D. wrote the manuscript. All authors contributed and agreed
- 317 on the contents.
- 318

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- 322 "Sciences de la Nature et de l'Homme", MNHN.
- 323

324 Notes:

- 325 Experimental procedures were carried out in accordance with European legislation on animal
- 326 experimentation (European Union Directive 2010/63/EU) and were approved for ethical
- 327 contentment by an independent ethical council (CEEA Cuvier n°68) and authorized by the
- 328 French government under reference number APAFiS#19316-2019032913284201 v1. Fish
- 329 were anesthetized in 0.1% tricaine methanesulfonate (MS-222; Sigma, St. Louis, MO)
- buffered with 0.1% NaHCO3 prior to sacrifice.
- 331

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- 336

337 Data availability

- 338 All R scripts are available on Github upon publication (https://github.com/PierreFoucault/).
- 339 All microbial communities' samples sequencing reads were deposited into the Sequence Read
- 340 Archive (SRA) database (accession number PRJNA836730 (samples SRR19170691 to
- 341 SRR19170767; Table S1).
- 342

343 **References:**

- 344 (1) Huisman, J.; Codd, G. A.; Paerl, H. W.; Ibelings, B. W.; Verspagen, J. M. H.; Visser,
- 345 P. M. Cyanobacterial Blooms. *Nat Rev Microbiol* **2018**, *16* (8), 471–483.
- 346 https://doi.org/10.1038/s41579-018-0040-1.
- 347 (2) Marie, B.; Huet, H.; Marie, A.; Djediat, C.; Puiseux-Dao, S.; Catherine, A.; Trinchet,
- I.; Edery, M. Effects of a Toxic Cyanobacterial Bloom (*Planktothrix agardhii*) on Fish:
- 349 Insights from Histopathological and Quantitative Proteomic Assessments Following the Oral
- 350 Exposure of Medaka Fish (Oryzias Latipes). *Aquatic Toxicology* **2012**, *114–115*, 39–48.

351 https://doi.org/10.1016/j.aquatox.2012.02.008. 352 (3)Saraf, S. R.; Frenkel, A.; Harke, M. J.; Jankowiak, J. G.; Gobler, C. J.; McElroy, A. E. 353 Effects of Microcystis on Development of Early Life Stage Japanese Medaka (Oryzias 354 latipes): Comparative Toxicity of Natural Blooms, Cultured Microcystis and Microcystin-LR. 355 Aquat Toxicol 2018, 194, 18–26. https://doi.org/10.1016/j.aquatox.2017.10.026. 356 (4) Le Manach, S.; Sotton, B.; Huet, H.; Duval, C.; Paris, A.; Marie, A.; Yépremian, C.; 357 Catherine, A.; Mathéron, L.; Vinh, J.; Edery, M.; Marie, B. Physiological Effects Caused by 358 Microcystin-Producing and Non-Microcystin Producing Microcystis aeruginosa on Medaka 359 Fish: A Proteomic and Metabolomic Study on Liver. Environ. Pollut. 2018, 234, 523–537. 360 https://doi.org/10.1016/j.envpol.2017.11.011. 361 Macke, E.; Callens, M.; De Meester, L.; Decaestecker, E. Host-Genotype Dependent (5) 362 Gut Microbiota Drives Zooplankton Tolerance to Toxic Cyanobacteria. Nat Commun 2017, 8 363 (1), 1608. https://doi.org/10.1038/s41467-017-01714-x. 364 McFall-Ngai, M.; Hadfield, M. G.; Bosch, T. C. G.; Carey, H. V.; Domazet-Lošo, T.; (6)365 Douglas, A. E.; Dubilier, N.; Eberl, G.; Fukami, T.; Gilbert, S. F.; Hentschel, U.; King, N.; Kjelleberg, S.; Knoll, A. H.; Kremer, N.; Mazmanian, S. K.; Metcalf, J. L.; Nealson, K.; 366 367 Pierce, N. E.; Rawls, J. F.; Reid, A.; Ruby, E. G.; Rumpho, M.; Sanders, J. G.; Tautz, D.; 368 Wernegreen, J. J. Animals in a Bacterial World, a New Imperative for the Life Sciences. 369 PNAS 2013, 110 (9), 3229–3236. https://doi.org/10.1073/pnas.1218525110. 370 Evariste, L.; Barret, M.; Mottier, A.; Mouchet, F.; Gauthier, L.; Pinelli, E. Gut (7)371 Microbiota of Aquatic Organisms: A Key Endpoint for Ecotoxicological Studies. Environ 372 Pollut 2019, 248, 989–999. https://doi.org/10.1016/j.envpol.2019.02.101. 373 Feng, P.; Xiao, X.; Zhou, T.; Li, X. Effects of the Bio-Accumulative Environmental (8) 374 Pollutants on the Gut Microbiota. In Gut Remediation of Environmental Pollutants; Li, X., 375 Liu, P., Eds.; Springer Singapore: Singapore, **2020**; pp 109–143. https://doi.org/10.1007/978-376 981-15-4759-1 4. 377 (9) Adamovsky, O.; Buerger, A. N.; Vespalcova, H.; Sohag, S. R.; Hanlon, A. T.; Ginn, P. 378 E.; Craft, S. L.; Smatana, S.; Budinska, E.; Persico, M.; Bisesi, J. H.; Martyniuk, C. J. 379 Evaluation of Microbiome-Host Relationships in the Zebrafish Gastrointestinal System 380 Reveals Adaptive Immunity Is a Target of Bis(2-Ethylhexyl) Phthalate (DEHP) Exposure. 381 Environ. Sci. Technol. 2020, 54 (9), 5719–5728. https://doi.org/10.1021/acs.est.0c00628. 382 (10)Duperron, S.; Halary, S.; Gallet, A.; Marie, B. Microbiome-Aware Ecotoxicology of 383 Organisms: Relevance, Pitfalls, and Challenges. Front. Public Health 2020, 8. 384 https://doi.org/10.3389/fpubh.2020.00407. 385 (11)Duperron, S.; Halary, S.; Habiballah, M.; Gallet, A.; Huet, H.; Duval, C.; Bernard, C.; 386 Marie, B. Response of Fish Gut Microbiota to Toxin-Containing Cyanobacterial Extracts: A 387 Microcosm Study on the Medaka (Oryzias latipes). Environ. Sci. Technol. Lett. 2019, 6 (6), 388 341–347. https://doi.org/10.1021/acs.estlett.9b00297. 389 Qian, H.; Zhang, M.; Liu, G.; Lu, T.; Sun, L.; Pan, X. Effects of Different (12)390 Concentrations of *Microcystis aeruginosa* on the Intestinal Microbiota and Immunity of 391 Zebrafish (Danio rerio). Chemosphere 2019, 214, 579–586. 392 https://doi.org/10.1016/j.chemosphere.2018.09.156. 393 (13)Ding, W.; Shangguan, Y.; Zhu, Y.; Sultan, Y.; Feng, Y.; Zhang, B.; Liu, Y.; Ma, J.; 394 Li, X. Negative Impacts of Microcystin-LR and Glyphosate on Zebrafish Intestine: Linked 395 with Gut Microbiota and MicroRNAs? Environ Pollut 2021, 286, 117685. 396 https://doi.org/10.1016/j.envpol.2021.117685. Gallet, A.; Halary, S.; Duval, C.; Huet, H.; Duperron, S.; Marie, B. Disruption of Fish 397 (14)398 Gut Microbiota Composition and Holobiont's Metabolome by Cyanobacterial Blooms. 399 *BioRxiv preprint* **2021**. https://doi.org/10.1101/2021.09.08.459397. 400 (15)Halary, S.; Duval, C.; Gallet, A.; Duperron, S.; Piquet, B.; Demay, J.; Bernard, C.;

- 401 Marie, B. Draft Genome Sequence of the Toxic Freshwater Microcystis aeruginosa Strain
- 402 PMC 728.11 (Cyanobacteria, Chroococcales). *Microbiol Resour Announc* **2020**, 9 (48).
- 403 https://doi.org/10.1128/MRA.01096-20.
- 404 (16) Hamlaoui, S.; Yéprémian, C.; Duval, C.; Marie, B.; Djédiat, C.; Piquet, B.; Bernard,
- 405 C.; Duperron, S. The Culture Collection of Cyanobacteria and Microalgae at the French
- 406 National Museum of Natural History: A Century Old But Still Alive and Kicking! Including
- 407 in Memoriam: Professor Alain Couté. *crya* **2022**, *43* (3), 41–83.
- 408 https://doi.org/10.5252/cryptogamie-algologie2022v43a3.
- 409 (17) Yéprémian, C.; Catherine, A.; Bernard, C.; Congestri, R.; Elersek, T.; Pilkaityte, R.
- 410 Chlorophyll a Extraction and Determination. In Handbook of Cyanobacterial Monitoring and
- 411 Cyanotoxin Analysis; John Wiley & Sons, Ltd, 2016; pp 331–334.
- 412 https://doi.org/10.1002/9781119068761.ch34.
- 413 (18) Rohart, F.; Gautier, B.; Singh, A.; Lê Cao, K.-A. MixOmics: An R Package for 'omics
- 414 Feature Selection and Multiple Data Integration. *PLoS Comput Biol* 2017, *13* (11), e1005752.
 415 https://doi.org/10.1371/journal.pcbi.1005752.
- 416 (19) Parada, A. E.; Needham, D. M.; Fuhrman, J. A. Every Base Matters: Assessing Small
- 417 Subunit RRNA Primers for Marine Microbiomes with Mock Communities, Time Series and
- 418 Global Field Samples. *Environ Microbiol* **2016**, *18* (5), 1403–1414.
- 419 https://doi.org/10.1111/1462-2920.13023.
- 420 (20) Bolyen, E.; Rideout, J. R.; Dillon, M. R.; Bokulich, N. A.; Abnet, C. C.; Al-Ghalith,
- 421 G. A.; Alexander, H.; Alm, E. J.; Arumugam, M.; Asnicar, F.; Bai, Y.; Bisanz, J. E.;
- 422 Bittinger, K.; Brejnrod, A.; Brislawn, C. J.; Brown, C. T.; Callahan, B. J.; Caraballo-
- 423 Rodríguez, A. M.; Chase, J.; Cope, E. K.; Da Silva, R.; Diener, C.; Dorrestein, P. C.; Douglas,
- 424 G. M.; Durall, D. M.; Duvallet, C.; Edwardson, C. F.; Ernst, M.; Estaki, M.; Fouquier, J.;
- 425 Gauglitz, J. M.; Gibbons, S. M.; Gibson, D. L.; Gonzalez, A.; Gorlick, K.; Guo, J.; Hillmann,
- 426 B.; Holmes, S.; Holste, H.; Huttenhower, C.; Huttley, G. A.; Janssen, S.; Jarmusch, A. K.;
- 427 Jiang, L.; Kaehler, B. D.; Kang, K. B.; Keefe, C. R.; Keim, P.; Kelley, S. T.; Knights, D.;
- 428 Koester, I.; Kosciolek, T.; Kreps, J.; Langille, M. G. I.; Lee, J.; Ley, R.; Liu, Y.-X.; Loftfield,
- 429 E.; Lozupone, C.; Maher, M.; Marotz, C.; Martin, B. D.; McDonald, D.; McIver, L. J.;
- 430 Melnik, A. V.; Metcalf, J. L.; Morgan, S. C.; Morton, J. T.; Naimey, A. T.; Navas-Molina, J.
- 431 A.; Nothias, L. F.; Orchanian, S. B.; Pearson, T.; Peoples, S. L.; Petras, D.; Preuss, M. L.;
- 432 Pruesse, E.; Rasmussen, L. B.; Rivers, A.; Robeson, M. S.; Rosenthal, P.; Segata, N.; Shaffer,
- 433 M.; Shiffer, A.; Sinha, R.; Song, S. J.; Spear, J. R.; Swafford, A. D.; Thompson, L. R.; Torres,
- 434 P. J.; Trinh, P.; Tripathi, A.; Turnbaugh, P. J.; Ul-Hasan, S.; van der Hooft, J. J. J.; Vargas, F.;
- 435 Vázquez-Baeza, Y.; Vogtmann, E.; von Hippel, M.; Walters, W.; Wan, Y.; Wang, M.;
- 436 Warren, J.; Weber, K. C.; Williamson, C. H. D.; Willis, A. D.; Xu, Z. Z.; Zaneveld, J. R.;
- 437 Zhang, Y.; Zhu, Q.; Knight, R.; Caporaso, J. G. Reproducible, Interactive, Scalable and
- 438 Extensible Microbiome Data Science Using QIIME 2. Nat Biotechnol 2019, 37 (8), 852–857.
- 439 https://doi.org/10.1038/s41587-019-0209-9.
- 440 (21) McMurdie, P. J.; Holmes, S. Phyloseq: An R Package for Reproducible Interactive
- 441 Analysis and Graphics of Microbiome Census Data. *PLoS One* **2013**, *8* (4), e61217.
- 442 https://doi.org/10.1371/journal.pone.0061217.
- 443 (22) Oksanen, J.; Simpson, G. L.; Blanchet, F. G.; Kindt, R.; Legendre, P.; Minchin, P. R.;
- 444 O'Hara, R. B.; Solymos, P.; Stevens, M. H. H.; Szoecs, E.; Wagner, H.; Barbour, M.;
- 445 Bedward, M.; Bolker, B.; Borcard, D.; Carvalho, G.; Chirico, M.; Caceres, M. D.; Durand, S.;
- 446 Evangelista, H. B. A.; FitzJohn, R.; Friendly, M.; Furneaux, B.; Hannigan, G.; Hill, M. O.;
- 447 Lahti, L.; McGlinn, D.; Ouellette, M.-H.; Cunha, E. R.; Smith, T.; Stier, A.; Braak, C. J. F. T.;
- 448 Weedon, J. Vegan: Community Ecology Package; **2022**.
- 449 (23) Hervé, M. RVAideMemoire: Testing and Plotting Procedures for Biostatistics; 2022.
- 450 (24) Tai, Y. C.; Speed, T. P. A Multivariate Empirical Bayes Statistic for Replicated

- 451 Microarray Time Course Data. *The Annals of Statistics* **2006**, *34* (5), 2387–2412.
- 452 (25) Huang, X.; Zeng, J.; Zhou, L.; Hu, C.; Yin, P.; Lin, X. A New Strategy for Analyzing
- 453 Time-Series Data Using Dynamic Networks: Identifying Prospective Biomarkers of
- 454 Hepatocellular Carcinoma. *Sci Rep* **2016**, *6* (1), 32448. https://doi.org/10.1038/srep32448.
- 455 (26) Singh, A.; Shannon, C. P.; Gautier, B.; Rohart, F.; Vacher, M.; Tebbutt, S. J.; Lê Cao,
- 456 K.-A. DIABLO: An Integrative Approach for Identifying Key Molecular Drivers from Multi-
- 457 Omics Assays. *Bioinformatics* **2019**, *35* (17), 3055–3062.
- 458 https://doi.org/10.1093/bioinformatics/bty1054.
- 459 (27) Wang, J.; Chen, W.-D.; Wang, Y.-D. The Relationship Between Gut Microbiota and
- 460 Inflammatory Diseases: The Role of Macrophages. *Frontiers in Microbiology* **2020**, *11*.
- 461 (28) Leitch, E. C. M.; Stewart, C. S. *Escherichia coli* O157 and Non-O157 Isolates Are
- 462 More Susceptible to l-Lactate than to d-Lactate. *Appl Environ Microbiol* **2002**, *68* (9), 4676– 463 4678. https://doi.org/10.1128/AEM.68.9.4676-4678.2002.
- 464 (29) Anderson, C. J.; Medina, C. B.; Barron, B. J.; Karvelyte, L.; Aaes, T. L.; Lambertz, I.;
- 465 Perry, J. S. A.; Mehrotra, P.; Gonçalves, A.; Lemeire, K.; Blancke, G.; Andries, V.; Ghazavi,
- 466 F.; Martens, A.; van Loo, G.; Vereecke, L.; Vandenabeele, P.; Ravichandran, K. S. Microbes
- 467 Exploit Death-Induced Nutrient Release by Gut Epithelial Cells. *Nature* 2021, 596 (7871),
 468 262–267. https://doi.org/10.1038/s41586-021-03785-9.
- 469 (30) Lee, Y.-S.; Kim, T.-Y.; Kim, Y.; Lee, S.-H.; Kim, S.; Kang, S. W.; Yang, J.-Y.; Baek,
- 470 I.-J.; Sung, Y. H.; Park, Y.-Y.; Hwang, S. W.; O, E.; Kim, K. S.; Liu, S.; Kamada, N.; Gao,
- 471 N.; Kweon, M.-N. Microbiota-Derived Lactate Accelerates Intestinal Stem-Cell-Mediated
- 472 Epithelial Development. *Cell Host Microbe* **2018**, *24* (6), 833-846.e6.
- 473 https://doi.org/10.1016/j.chom.2018.11.002.
- 474 (31) Guo, F.-F.; Yu, T.-C.; Hong, J.; Fang, J.-Y. Emerging Roles of Hydrogen Sulfide in
 475 Inflammatory and Neoplastic Colonic Diseases. *Frontiers in Physiology* 2016, 7.
- 476 (32) Portz, D. E.; Woodley, C. M.; Cech, J. J. Stress-Associated Impacts of Short-Term
- 477 Holding on Fishes. Rev Fish Biol Fisheries 2006, 16 (2), 125–170.
- 478 https://doi.org/10.1007/s11160-006-9012-z.
- 479 (33) Ernst, B.; Hitzfeld, B.; Dietrich, D. Presence of *Planktothrix sp.* and Cyanobacterial
 480 Toxins in Lake Ammersee, Germany and Their Impact on Whitefish (Coregonus Lavaretus)
- 481 L.). *Environmental Toxicology* **2001**, *16* (6), 483–488. https://doi.org/10.1002/tox.10006.
- 482 (34) Le Manach, S.; Khenfech, N.; Huet, H.; Qiao, Q.; Duval, C.; Marie, A.; Bolbach, G.;
- 483 Clodic, G.; Djediat, C.; Bernard, C.; Edery, M.; Marie, B. Gender-Specific Toxicological
- 484 Effects of Chronic Exposure to Pure Microcystin-LR or Complex *Microcystis aeruginosa*
- 485 Extracts on Adult Medaka Fish. *Environ Sci Technol* **2016**, *50* (15), 8324–8334.
- 486 https://doi.org/10.1021/acs.est.6b01903.
- 487 (35) Bridgewater, L. C.; Zhang, C.; Wu, Y.; Hu, W.; Zhang, Q.; Wang, J.; Li, S.; Zhao, L.
- 488 Gender-Based Differences in Host Behavior and Gut Microbiota Composition in Response to
- 489 High Fat Diet and Stress in a Mouse Model. Sci Rep 2017, 7 (1), 10776.
- 490 https://doi.org/10.1038/s41598-017-11069-4.
- 491

492 Supplementary information: supplementary Material and methods, Fig S1 and Fig S2

493 Supplementary tables are summarized in a single supplementary file containing the

- 494 **following:**
- 495 Table S1: Parameters monitoring from the microcosm experiment
- 496 Table S2: SRA accession numbers
- 497 Table S3: Sample IDs, raw reads counts, Sequencing depth and read filtering

- 498 Table S4: Gut microbiota samples median alpha-diversity indices and significativity group
- 499 Table S5: Gut microbiota samples alpha and beta diversity metrics; and gut, liver and muscles
- 500 group comparison scores and significance levels
- 501 Table S6: Gut microbiota phylum and genera MEBA Hoteling T², Kruskal Wallis and
- 502 Wilcoxon post-hoc significativity levels (ns: not significant; *<0.05; **<0.01; ***<0.001)
- 503 Tables S7, S8 and Table S9: Gut, liver and muscle metabolites log-normalized count tables